



## Can CO<sub>2</sub> sensors in the ventilation system of a pool facility help reduce the variability in the trihalomethane concentration observed in indoor air?



Therese B. Nitter<sup>a,\*</sup>, Morten S. Grande<sup>b</sup>, Kristin V.H. Svendsen<sup>c</sup>, Rikke B. Jørgensen<sup>c</sup>, Salvatore Carlucci<sup>d</sup>, Guangyu Cao<sup>b</sup>

<sup>a</sup> Department of Civil and Environmental Engineering, Norwegian University of Science and Technology (NTNU), Norway

<sup>b</sup> Department of Energy and Process Engineering, NTNU, Norway

<sup>c</sup> Department of Industrial Economics and Technology Management, NTNU, Norway

<sup>d</sup> Energy, Environment and Water Research Centre, The Cyprus Institute, Nicosia, Cyprus

### ARTICLE INFO

Handling Editor: Xavier Querol

#### Keywords:

Swimming facility  
Carbon dioxide  
Trihalomethane  
Ventilation  
Air quality  
Random intercept model

### ABSTRACT

Volatile and hazardous compounds are formed during the chlorination of pool water. Monitoring components in the air, such as the four trihalomethanes; chloroform, dichlorobromomethane, dibromochloromethane and bromoform (tTHM), is challenging. Carbon dioxide (CO<sub>2</sub>) sensors are used for controlling air quality in different buildings and can be installed in ventilation systems for continuous surveillance and monitoring purposes. However, such sensors are not used in indoor swimming facilities. In this study, samples of tTHM and CO<sub>2</sub> were collected and analysed, along with other air and water quality parameters such as combined chlorine, to evaluate whether CO<sub>2</sub> sensors could be used to explain the observed variability in the tTHM concentration in an indoor swimming facility and thereby reduce the exposure of individuals utilising the pool to tTHM. Random intercept models were built for the tTHM and CO<sub>2</sub> concentrations, respectively, and the results show that the relationships between combined chlorine in the water, CO<sub>2</sub> in the air and number of occupants explain 52% of the variability in tTHM. The correlation between occupancy and CO<sub>2</sub> concentration ( $\rho = 0.65$ ,  $p \leq 0.01$ ) suggests that CO<sub>2</sub> sensors should be used so that the air supply corresponds to the demand of the users.

### 1. Introduction

People in developed countries spend an average of 80–90% of their time indoors (ASHRAE, 2016), and sufficient indoor air quality (IAQ) is necessary to maintain a healthy indoor environment. However, due to the evaporation of potentially hazardous gasses from the pool water's surface, indoor swimming facilities introduce unique IAQ challenges compared to those observed in offices and residential buildings (Lebon et al., 2017). In pool facilities, there are increased risks of moisture damage, bacterial growth and corrosion (Ciuman and Lipska, 2018; Liu et al., 2018), and in many facilities, the occupants complain regularly about thermal comfort, respiratory irritations and skin problems (Nitter et al., 2019).

Although improper disinfection is associated with outbreaks of fatal contaminants in the water (World Health Organization, 2006), the reaction between free chlorine and precursors in the pool water, introduced by swimmers and from filling water (Deutsches Institut für Normung (DIN), 2012), also leads to the formation of disinfection by-

products (DBPs) (Daiber et al., 2016). Volatile DBPs, such as trihalomethanes (THM), can be potentially hazardous to human health if individuals are exposed to high concentrations over a long period of time (Font-Ribera et al., 2018; Gouveia et al., 2019). Typically the following four THM, referred to as total THM (tTHM) are formed as a result of chlorination: bromoform, dibromochloromethane, chloroform and dichlorobromomethane, of which the last two are characterized as potentially carcinogenic to humans by the International Agency for Research on Cancer (IARC) (World Health Organization, 2017). In previous studies, the air concentrations of tTHM have been found to correlate with the volatile trichloramine (NCl<sub>3</sub>) concentration (Cossec et al., 2016; Nitter and Svendsen, 2019a), which is related to the increased prevalence of respiratory irritations amongst swimmers and lifeguards (Chu et al., 2013; Jacobs et al., 2007; Lévesque et al., 2006; Andersson et al., 2018).

In order to control the relative humidity (RH) and air temperature in the poolroom, the room is ventilated mechanically. The air supply grills are typically located at floor level, and the air is supplied to the room at

*Abbreviations:* ACH, air changes per hour; AR(1), first order autoregressive; CO<sub>2</sub>, carbon dioxide; DBP, disinfection by-product; IAQ, indoor air quality; RH, relative humidity; tTHM, sum of the four most common trihalomethanes; UV, ultra violette

\* Corresponding author at: Høgskoleringen 7A, 7491 Trondheim, Norway.

E-mail address: [therese.nitter@ntnu.no](mailto:therese.nitter@ntnu.no) (T.B. Nitter).

<https://doi.org/10.1016/j.envint.2020.105665>

Received 31 January 2020; Received in revised form 12 March 2020; Accepted 13 March 2020

Available online 19 March 2020

0160-4120/© 2020 The Authors. Published by Elsevier Ltd. This is an open access article under the CC BY license (<http://creativecommons.org/licenses/by/4.0/>).

relatively high velocities to mix the air by jets up along the window façade. However, the distribution of air in such facilities is often complex and inadequate (Lebon et al., 2017; Nitter and Svendsen, 2019b). Worldwide, the most common measure for the concentration of precursors in swimming pool water is combined chlorine, a product of the reaction between ammonia and free chlorine. Combined chlorine is also the only DBP for which a pool water limit value exists in Norway; no other DBPs are being controlled either in the air or in the water (Norwegian Ministry of Health, 1996). The German Federal Environmental Agency suggests that the concentration of chloroform (one of the four THMs) in the air should never exceed  $200 \mu\text{g}/\text{m}^3$ , as this value indicates improper air and/or water quality (Verein Deutscher Ingenieure, 2010).

The absence of required limit values for DBPs in the air of pool facilities makes it difficult to judge whether the air renewal is adequate. Monitoring the most hazardous components is also challenging, as no easy measurement techniques, such as sensors for characterizing  $\text{NCl}_3$  and tTHM concentrations in the air, exist.

Carbon dioxide ( $\text{CO}_2$ ) is exhaled when people breathe, and this component is considered to be a good indicator of the number of occupants in a given room (Dougan and Damiano, 2004; Seppänen et al., 1999). The development of different sensors for the continuous surveillance and measurement of  $\text{CO}_2$  have also made it easy to monitor this gas (Norbäck et al., 1995; Qiao et al., 2019). In Norway, it is recommended that the air supply be controlled in sport halls with the use of sensors for  $\text{CO}_2$  and air temperature (Ministry of Culture, 2016). Such sensors, however, are not used in swimming facilities.

Findings from previous studies show that the concentration of  $\text{CO}_2$ , in addition to the occupancy level, could function as an indicator for other components that are related to illness (Norbäck et al., 1995; Padhi et al., 2012; Rodríguez et al., 2018). Studies have also identified insufficient ventilation, such as condensation on window surfaces, high  $\text{CO}_2$  concentrations and occupancy level, to be associated with high tTHM concentrations (Gabriel et al., 2019). Based on knowledge from previous studies, the aim of the present study is to investigate whether the measured  $\text{CO}_2$  concentration can be used as an effective indicator for predicting tTHM concentration in a swimming pool facility.

## 2. Materials and methods

### 2.1. Pool dimensions, air handling and water treatment

In this study, one poolroom built in 2018 and containing one swimming pool (12 m  $\times$  8 m) was investigated. The total air volume in the poolroom was approximately  $1050 \text{ m}^3$ . The swimming pool was filled with freshwater and disinfected using sodium hypochlorite ( $\text{NaOCl}$ ) in addition to UV treatments. During the sampling days, the air change rate (ACH) varied between 5.1 and  $5.7 \text{ h}^{-1}$ , with a fresh air rate of between 70 and 100%. During night-mode ventilation, the ACH was reduced to 60–70% of the day-mode ventilation. The air supply was controlled using set points of RH and air temperature and was preheated in the ventilation unit before being supplied to the room. The air was supplied to the room by grills located on the floor and up along the window façade. To mix the air in the room, the air was supplied at relatively high velocity. During air sampling, the swimming pool was being used mainly for swimming education.

### 2.2. Sampling strategy

Due to the sampling and analytical procedures, each tTHM sample had to be collected over 20 min. In total, 65 samples of tTHM were collected, simultaneously with  $\text{CO}_2$ , RH and air temperature, over the course of three Tuesdays and four Thursdays during a four-week period.  $\text{CO}_2$ , RH and air temperature were collected at intervals of two minutes and were logged continuously while present in the poolroom. One 20-minute sample of tTHM was collected every 30 min from 10:00 to 15:00, except for the final day, when samples were collected from 12:00

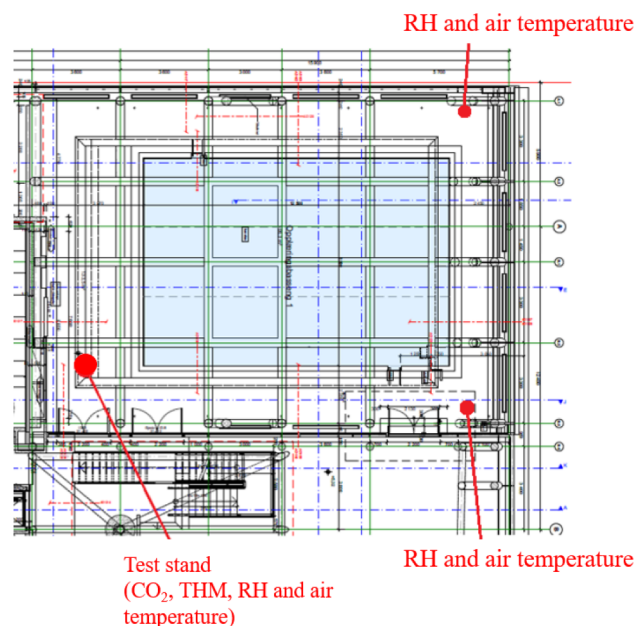


Fig. 1. Sketch of the pool facility with its sampling locations.

to 14:00. All air quality parameters (tTHM,  $\text{CO}_2$ , air temperature and RH) were collected from 0.30 m above the floor level using a test stand. The location of the test stand is shown in Fig. 1. In a previous study, no differences in tTHM concentration were found between 0.05 m and 0.30 m, and the concentrations measured from 0.30 m were therefore assumed to be representative of the concentration in the breathing zone of the occupants in the pool (Nitter and Svendsen, 2019a). The test stand was placed approximately 2.8 m from the pool edge to prevent interference with the activity in the room. Before sampling was carried out, the air temperature and RH were measured from two different locations in the room to investigate whether the air in the room could be considered fully mixed. The two locations are shown in Fig. 1.

### 2.3. Air samples and analysis

Samples were collected using the active air sampling method, by which air is pulled into an automatic thermal desorption (ATD) tube using a pump. The ATD tubes contained 200 mg of Tenax TA 35/60 (Markes Int, 2019b) and were fastened to one Acti-Voc low-flow sampling pump (Markes Int.) (Markes Int, 2019a). The pump was calibrated to deliver a flow rate of 40 ml/min for 20 min, which has been found to provide satisfactory results with regards to both safe sampling volume and uncertainty (Nitter and Svendsen, 2019b). The pump was calibrated in the poolroom before and after each sample. The ATD tubes were sealed with Swagelok fittings with PTFE ferrules and packed in uncoated aluminium foil, both before leaving the lab and immediately after sampling. To prevent the ATD tubes from being contaminated, they were always stored in an airtight container with charcoal when not in use. The sampling, analysis, and quality assurance for collecting samples of tTHM in the air were based on the methods published in US EPA TO-17 (United States Environmental Protection Agency, 1999) and the ISO standard 16017 (International Organization for Standardization, 2000). Samples of tTHM were identified and quantified using a GC/MSD in the laboratory of Health Safety and Environment at the Department of Industrial Economics and Technology Management, NTNU, Norway. The analysis setup is explained elsewhere (Nitter et al., 2018).

The air temperature and RH were recorded every two minutes using one EasyLog USB (EL-USB-2). The concentration of  $\text{CO}_2$  was measured every second minute using a KIMO AQ 200. Information on tube number, time, flow rate, water activity and number of occupants was logged for each sample of tTHM collected. The water concentration of

**Table 1**

Mean for physical and chemical parameters and number of occupants for the different sampling days.

Day	Number of Occupants	Cl <sub>Comb</sub> (mg/l)	T <sub>air</sub> (°C)	RH (%)	tTHM (µg/m <sup>3</sup> )	CO <sub>2</sub> (ppm)	ACH (h <sup>-1</sup> )	n <sup>†</sup>
1	7	0.12	29.0		112.0	626	5.7	10
2	14	0.15	29.0		116.3	641	5.1	10
3	14	0.13	28.4	56.9	116.8	639	5.6	10
4	15	0.10	28.6	55.0	88.9	649	5.7	10
5	13	0.19	28.5	60.0	134.6	668	5.5	10
6	5	0.20	28.7	56.1	126.8	620	5.6	10
7	14	0.20	29.0		184.0	659	5.1	5

Abbreviations: Cl<sub>Comb</sub> stands for combined chlorine, T<sub>air</sub> stands for air temperature.

<sup>†</sup> Based on the number of tTHM samples collected.

free and combined chlorine was logged continuously in the facility. To calibrate the logging instrument, the concentrations of free and combined chlorine were measured manually at least three times per day during open hours, in accordance with the Norwegian regulations (Norwegian Ministry of Health, 1996). After each day of sampling, the concentrations of free and combined chlorine, pH and water temperature were provided by the technical staff who performed these measurements.

#### 2.4. Statistical analysis

All variables observed in this study were interpreted and analysed using the statistics software package STATA 15.1. The average measured CO<sub>2</sub> concentration, RH and air temperature for each tTHM sample (lasting for 20 min) was used in the statistical analysis. Descriptive data for the different variables measured in this study are presented in Table 1. Due to some skewness in the data, tTHM and CO<sub>2</sub> concentrations were transformed using the natural logarithm function before parametric methods were applied. The residuals for the tTHM and CO<sub>2</sub> concentrations were tested for normality using the Shapiro-Wilk test and plotted using histograms. Based on the results, the null hypothesis of normality could not be rejected.

The dataset consisted of both continuous variables (such as tTHM and CO<sub>2</sub> concentrations, number of occupants, air temperature, RH, ACH and combined chlorine) and ordinal cyclic variables (such as day of the week and time during the day). To estimate the independence and variability of the collected data, random intercept models for tTHM and CO<sub>2</sub> concentrations were built, where day was used as cluster unit, and time was used as unit for repeated measures. In the random intercept model, the natural logarithm transformations of the two components (ln C<sub>ijk</sub>) were used as dependent variables. In the model for tTHM concentration, the number of occupants, the concentration of combined chlorine and the CO<sub>2</sub> concentration were interpreted as fixed effects. The same approach was used for CO<sub>2</sub>, but, in this second model,

**Table 2**

Significant Spearman's correlations (ρ) with tTHM and CO<sub>2</sub> concentrations.

	Number of Occupants	Cl <sub>comb</sub>	tTHM	CO <sub>2</sub>	RH	ACH	T <sub>air</sub>
Number of Occupants							
Cl <sub>comb</sub>	-0.11						
THM	0.22	0.61**					
CO <sub>2</sub>	0.65**	0.02	0.34**				
RH	0.25	0.22	0.53**	0.46**			
ACH	-0.03	-0.44**	-0.33**	0.02	-0.18		
T <sub>air</sub>	-0.26*	-0.00	-0.1	-0.25*	-0.35*	-0.14	

Abbreviations: T<sub>air</sub> stands for air temperature, Cl<sub>comb</sub> stands for combined chlorine.

\* p ≤ 0.05.

\*\* p ≤ 0.01.

the tTHM concentration was included as a fixed effect, and the concentration of combined chlorine was excluded, as this variable did not explain any of the observed variability. The geometric mean exposure, as determined by the identified variables, can be estimated by back transformation of the regression coefficients using the following formula:

$$E = e^{c+b_1+b_2\cdots} = e^c \times effect_{determinant_1} \times effect_{determinant_2} \quad (1)$$

where  $E$  is the exposure,  $c$  is the intercept of the regression model and  $b_1$  and  $b_2$  are the regression coefficients of the predictor variables. The final model was estimated using the method of restricted maximum likelihood. Other variables, such as the interaction between combined chlorine and number of occupants, RH, air temperature and ACH, were tested for both models, but, as they were not statistically significant (at  $p \leq 0.05$ ), they were taken out of the model again. IAQ parameters are often autocorrelated as a result of limited mixing of contaminants and insufficient air exchange (Luoma and Batterman, 2000). To account for the potential correlation between the repeated samples collected on the same day, different covariance structures were tested using the log likelihood ratio test. The covariance first-order autoregressive (AR (1)) structure was used for both models. The AR (1) structure assumes that the correlation function decays exponentially as the intervals between the measurements increase (Peretz et al., 2002).

### 3. Results

In Table 1, an overview of the mean values for the different variables measured in this study is shown. The water temperature, concentration of free chlorine and pH value were stable throughout the study period at 31 °C, 0.8 mg/l and 7.2, respectively; therefore, these values are not included in the table. As shown, the average daily concentrations of tTHM and CO<sub>2</sub> measured ranged from 88.9 µg/m<sup>3</sup> to 184.0 µg/m<sup>3</sup> and 620 ppm to 668 ppm respectively. Except for dibromochloromethane, which was not quantifiable on three samples, all four tTHM were quantified in all samples, in which chloroform accounted for 82% of the quantified tTHM, while bromodichloromethane, dibromochloromethane and bromoform accounted for 9.5%, 1.0% and 7.5% respectively. During the period of sampling, the mean outdoor air temperature, measured from the city weather station, varied between -9.2 °C and 7.7 °C. The measured air temperature was very stable, and the difference between the lowest and highest measure was only 0.6 °C.

In Table 2, the parameters that were significantly correlated with CO<sub>2</sub> and THM concentrations are shown. As expected, a statistically significant correlation between the measured level of CO<sub>2</sub> and number of occupants in the room was obtained ( $\rho = 0.645$ ,  $p = 0.01$ ). A significant Pearson's correlation was also obtained between the natural logarithmically transformed tTHM concentration and the natural logarithmically transformed CO<sub>2</sub> concentration ( $r = 0.38$ ,  $p \leq 0.01$ ). Both the CO<sub>2</sub> and tTHM concentrations are significantly and positively correlated with RH; i.e. when RH increases, air contamination rises. A significant negative correlation between tTHM concentration and ACH was also found.

**Table 3**  
Random intercept model for Ln tTHM with estimates of random effects.

Parameter <sup>†</sup>	Estimate <sup>‡</sup>	95% Confidence interval	
Intercept	3.036	2.34	3.73
Number of occupants	0.005	0.00	0.01
Combined chlorine	4.299	0.91	7.69
CO <sub>2</sub> concentration	0.002	0.001	0.003
Random effects <sup>†</sup>	Variance explained by random effects		
Within day variability ( $\sigma_w^2$ )	0.021	52%	
Between day variability ( $\sigma_b^2$ )	0.012		
Correlation between repeated measures	0.73		

<sup>†</sup> All parameters are significant at  $p < 0.05$ .

<sup>‡</sup> The estimates are for each one unit increase in the parameters and how much this increases the tTHM concentration. Example: One ppm increase in CO<sub>2</sub> gives an  $e^{0.002} = 1.002$  or a 0.2% increase in tTHM. The estimated model parameters may only be valid for the observed values.

The parameter estimates for tTHM concentration using a random intercept model are shown in Table 3. Before the variables were added to the model, the estimated total variability was 0.070. After the three significant variables (i.e. CO<sub>2</sub> concentration, number of occupants and concentration of combined chlorine in the water) were included in the model, the total variability was reduced to 0.033, meaning that three variables thereby explained 52% of the total variability observed. The correlation between the repeated observations within the same day was estimated as 0.73, meaning that ignoring the correlation between the repeated observations might lead to incorrect parameter estimates.

In Table 4, the random intercept model for the CO<sub>2</sub> concentration is shown. For this component, only two explanatory variables (tTHM concentration and number of occupants) contributed significantly to the model. These two variables explained 44% of the observed variability. The correlation between the repeated observations was low (0.081), meaning that the observations of CO<sub>2</sub> concentrations can be assumed to be approximately independent of these variables and that other independent observations might be useful in analysing the results.

The CO<sub>2</sub> concentration can be estimated using Eq. (1) and the regression coefficient in Table 4. For example, if the concentration of tTHM is 200  $\mu\text{g}/\text{m}^3$  and the number of occupants is 20, results in an estimated geometric mean CO<sub>2</sub> concentration of:

$$E = e^{c+b_1+b_2\cdots} = e^{6.3365} \times e^{(0.00426 \times 20)} \times e^{(0.00063 \times 200)} = 698 \text{ ppm}$$

#### 4. Discussion

Epidemiological evidence suggests that there is an association between exposure in swimming pool facilities and health effects such as irritations to the skin, eyes and respiratory tract and even cancer (Gouveia et al., 2019; Fantuzzi et al., 2010; Hery et al., 1995; Jacobs et al., 2007). To

**Table 4**  
Random intercept model for Ln CO<sub>2</sub> with estimates of random effects.

Parameter <sup>†</sup>	Estimate	95% Confidence interval	
Intercept	6.337	6.28	6.40
Number of occupants	0.004	0.003	0.006
tTHM	0.001	0.0001	0.001
Random effects <sup>†</sup>	Variance explained by random effects		
Within day variability ( $\sigma_w^2$ )	0.003	44%	
Between day variability ( $\sigma_b^2$ )	0.000		
Correlation between repeated measures	0.081		

<sup>†</sup> All parameters are significant at  $p < 0.05$ .

protect people who are regularly exposed in such environments from increased risk of disease, the implementation of guidelines and control strategies is considered necessary. Volatile compounds, such as NCl<sub>3</sub> and tTHM, in the air of indoor swimming facilities have been studied in previous literature (Afifi and Blatchley, 2015; Hsu et al., 2009; Nitter and Svendsen, 2019a); however, these components are difficult and expensive to measure and analyse, and no sensor technology allowing for continuous monitoring exists. Air concentrations of tTHM may be used as an indicator for the air concentration of NCl<sub>3</sub> (Nitter and Svendsen, 2019a; Cosset et al., 2016). If sensors for CO<sub>2</sub> could be used to predict the tTHM concentrations in the air, controlling the air quality in swimming facilities would become less complex. In this study, the indoor concentration of CO<sub>2</sub> was measured in parallel with the tTHM concentration in order to investigate whether CO<sub>2</sub> could be used to estimate the number of occupants and function as an indirect indicator for the tTHM concentration in the air.

##### 4.1. Can CO<sub>2</sub> be used to estimate contamination by tTHM?

The highest value of tTHM, 184  $\mu\text{g}/\text{m}^3$ , was measured the final day of sampling and is close to the threshold value of 200  $\mu\text{g}/\text{m}^3$  recommended by the German Federal Environmental Agency (Verein Deutscher Ingenieure, 2010). In a recent study, where mean air concentrations of tTHM was measured to be 205  $\mu\text{g}/\text{m}^3$ , the cancer risk among elite swimmers was found to be unacceptably high (Gouveia et al., 2019), and therefore keeping the concentrations below 200  $\mu\text{g}/\text{m}^3$  is considered necessary to protect the occupants in the poolroom. Pearson's correlation between the air concentrations of CO<sub>2</sub> and tTHM shows a statistically significant relationship between the two ( $r = 0.38$ ,  $p \leq 0.01$ ). However, this relationship is far from linear, and the CO<sub>2</sub> concentration is not considered an optimal means for controlling the air concentration of tTHM. As shown in Table 3, the concentration of tTHM depends on the water concentration of combined chlorine, occupancy level and CO<sub>2</sub>, and these variables combined explained 52% of the observed variability in tTHM concentration. If one assumes that the concentration of combined chlorine in the water is stable around 0.15 mg/l and that the pool capacity is 20 people, then one can allow the CO<sub>2</sub> concentration to be around 700 ppm while keeping the geometric mean concentration of tTHM below 200  $\mu\text{g}/\text{m}^3$ . This limit value may only help control the IAQ in poolrooms of the same size, same water quality, same occupancy level and same ventilation strategy as in the investigated poolroom.

As of today, the air supply in this facility is controlled using sensors for RH and air temperature. Controlling these parameters is necessary to control the energy use and to protect the building construction (SINTEF Byggforsk, 2003), and these sensors allow the amount of recirculated air to be adjusted. The dehumidification unit makes the system robust towards changes in outdoor and indoor conditions concerning RH. However, as shown in Table 2, a negative correlation between ACH and tTHM was observed, meaning that when the ACH increases, the tTHM concentration decreases. Adjusting the amount of recirculated air or ACH to adjust RH will therefore cause changes in the level of tTHM or other DBPs. Overall, little variation of CO<sub>2</sub> and tTHM were observed, which is likely to be a result of the high ACH and fresh air supply used in this poolroom. The investigated poolroom is also small, consisting of only one swimming pool, and the chosen ventilation strategy is assumed to be effective (Nitter and Svendsen, 2019a). In larger pool facilities, however, the ventilation efficiency might not be considered equally good (Nitter and Svendsen, 2019b), and lower ACH, which is related to the accumulation of tTHM in the air and an increase in reported health issues, is common (Bessonneau et al., 2011; Nitter et al., 2019).

tTHM is a product of the reaction between precursors from the number of occupants in the pool, precursors in the filling water and chlorine, and the formation rate depends on the water circulation system, the disinfecting strategy, water temperature, pH value and the concentration of bromine (World Health Organization, 2017). While CO<sub>2</sub> is generated by the occupants in the room and thereby will be reduced to outdoor concentrations when no occupants are present (Dougan and

Damiano, 2004), tTHM will be transported from the water to the air even after the occupants have left the swimming pool. This fact also explains the large correlation obtained between the repeated measures of tTHM (see Table 3), as the concentration levels are not only dependent upon the occupancy level, as is the case for the concentration of CO<sub>2</sub>. Despite the complex nature of swimming facilities, with their varying sizes, ceiling heights and different user groups, a minimum requirement for ACH and the implementation of CO<sub>2</sub> sensors are assumed to make the system more robust towards sudden changes in occupancy or activity level as well as reducing the observed variability in tTHM.

#### 4.2. Can CO<sub>2</sub> be used to predict the number of occupants in a poolroom?

In a previous study that used a cross-sectional study design, CO<sub>2</sub> concentrations were measured in the air of 20 swimming facilities, and these concentrations were found to vary significantly between the facilities, ranging from 351 ppm to 1553 ppm (Gabriel et al., 2019). In this study, the CO<sub>2</sub> concentrations in the poolroom investigated decreased quickly after each swimming session and when no people were present, which indicates that the ventilation system replaced the air in the room effectively (Lu et al., 2015). As shown in Table 4, effective air exchange is also confirmed by the lack of correlation observed between the repeated measurements of CO<sub>2</sub> (Luoma and Batterman, 2000).

The main predictors for energy consumption in a swimming facility are floor area, surface of the swimming pool and number of visitors (Kampel et al., 2016; Nitter et al., 2019). During the days of sampling, the bather load varied from 0 to 30 people in the pool. While the ACH is controlled after design criteria and therefore varies little over the day, the bather load varies significantly. Considering bather load to be one of the main predictors for energy consumption, creating a more dynamic system corresponding to the user demand can potentially reduce the energy use as well as improve the air quality.

As shown in Table 2, a statistically significant correlation was obtained between CO<sub>2</sub> concentration and the number of occupants ( $\rho = 0.654, p \leq 0.01$ ), which corresponds to the correlation between the number of occupants and CO<sub>2</sub> concentration found in previous studies (Gabriel et al., 2019; Lu et al., 2015; Rodríguez et al., 2018). The results obtained in this study might indicate that the air supply should be controlled with respect to the CO<sub>2</sub> level so that more fresh air could be distributed to the poolroom during periods of occupancy. However, such a tactic might only be suitable for rooms where the air can be considered well mixed, which might not be the case for larger pool facilities and water parks. It also requires having a relatively high ACH for the sensors to detect concentrations of CO<sub>2</sub> representative of that in the users' breathing zone. For buildings where the ACH is low (0.2–0.5 h<sup>-1</sup>), there might be a delay between the response from the ventilation system and supply of fresh air, which, in some cases, might result in fresh air being supplied to the room after the occupants have left the location. In previous studies, correlations have been found between the number of occupants and the respiratory irritant NCl<sub>3</sub>, and between tTHM and NCl<sub>3</sub> in the air (Nitter and Svendsen, 2019a; Cossec et al., 2016). Adjusting the air supply based on the number of occupants using CO<sub>2</sub> sensors might therefore also control the concentration of NCl<sub>3</sub>. This assumption, however, should be investigated further before any conclusions are made.

When CO<sub>2</sub> concentration is used as an indicator to predict the number of occupants in a room, then the underlying assumption is that the occupants have the same metabolic rate, diet and activity level (Dougan and Damiano, 2004). A poolroom, however, is typically used by individuals in different age groups, for different purposes and with dissimilar metabolic rates. Therefore, individuals might differ significantly in terms of the release of CO<sub>2</sub>. Controlling the supply airflow rate based on the CO<sub>2</sub> level will increase the fresh air supply when the need for fresh air increases, regardless of the level of occupancy. The integration of CO<sub>2</sub> sensors into the ventilation system might make the ventilation strategy more dynamic and better able to correspond to visitor and activity level patterns.

## 5. Conclusions

The aim of this study was to investigate whether the measured CO<sub>2</sub> concentration can be used as an effective indicator for predicting tTHM concentration in a swimming pool facility. The results show that the CO<sub>2</sub> concentration alone may not function as an optimal indicator for predicting the air concentration of tTHM. Rather, the CO<sub>2</sub> concentration, in combination with the occupancy level and water concentration of combined chlorine can improve the control of the air exposure to tTHM in this swimming facility and these predictor variables explained 52% of the variability observed in tTHM. The correlation between occupancy level and CO<sub>2</sub> ( $\rho = 0.65, p \leq 0.01$ ) also suggests that CO<sub>2</sub> sensors should be used to increase the air supply during occupancy and reduce the air supply during non-occupancy periods to save energy. A significant negative correlation between ACH and tTHM was obtained, and a minimum requirement of ACH and fresh air supply should be implemented to prevent tTHM to accumulate in the air.

## Funding

This research did not receive any specific grant from funding agencies in the public, commercial, or not-for-profit sectors.

## CRedit authorship contribution statement

**Therese B. Nitter:** Conceptualization, Methodology, Software, Validation, Formal analysis, Investigation, Resources, Writing - review & editing, Visualization, Project administration. **Morten S. Grande:** Conceptualization, Investigation, Writing - review & editing, Project administration. **Kristin V.H. Svendsen:** Conceptualization, Methodology, Validation, Resources, Writing - original draft, Writing - review & editing, Supervision. **Rikke B. Jørgensen:** Conceptualization, Resources, Writing - original draft, Writing - review & editing, Supervision. **Salvatore Carlucci:** Conceptualization, Writing - original draft, Writing - review & editing, Supervision. **Guangyu Cao:** Conceptualization, Methodology, Writing - review & editing, Supervision.

## Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

## Acknowledgements

The authors would also like to thank the swimming facility office for collaborating and providing necessary information. We would also like to thank Bjørn Aas at the Centre for Sports Facilities and Technology (SIAT), located in the Department of Civil and Environmental Engineering at the Norwegian University of Science and Technology (NTNU), for proofreading.

## Appendix A. Supplementary material

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.envint.2020.105665>.

## References

- Afifi, M.Z., Blatchley, E.R., 2015. Seasonal dynamics of water and air chemistry in an indoor chlorinated swimming pool. *Water Res.* 68, 771–783. <https://doi.org/10.1016/j.watres.2014.10.037>.
- ASHRAE, 2016. I Interactions Affecting the Achievement of Acceptable Indoor Environments. Guideline 10.
- Bessonneau, V., Derbez, M., Clément, M., Olivier, T., 2011. Determinants of chlorination by-products in indoor swimming pools. *Int. J. Hyg. Environ. Health* 215 (1), 76–85. <https://doi.org/10.1016/j.ijheh.2011.07.009>.

- Peretz, C., Goren, A., Smid, T., Kromhout, H., 2002. Application of mixed-effects models for exposure assessment. *Ann. Work Expo Health* 46 (1), 69–77. <https://doi.org/10.1093/annhyg/mef009>.
- Cossec, C.L., Laurent, A.-M., Person, A., Rouvié-Laurie, I., Beaubestres, C., 2016. Trichloramine and trihalomethanes concentrations in air or water of Paris indoor swimming pools and impact of different water treatment methods. *Pollut. Atmos.* 228, 1–14.
- Chu, T.-S., Cheng, S.-F., Wang, G.-S., Tsai, S.-W., 2013. Occupational exposures of airborne trichloramine at indoor swimming pools in Taipei. *Sci. Total Environ.* 461–462, 317–322. <https://doi.org/10.1016/j.scitotenv.2013.05.012>.
- Ciunan, P., Lipska, B., 2018. Experimental validation of the numerical model of air, heat and moisture flow in an indoor swimming pool. *Build. Environ.* 145, 1–13. <https://doi.org/10.1016/j.buildenv.2018.09.009>.
- Daiber, E.J., DeMarini, D.M., Ravuri, S.A., et al., 2016. Progressive increase in disinfection byproducts and mutagenicity from source to tap to swimming pool and spa water: impact of human inputs. *Environ. Sci. Technol.* 50 (13), 6652–6662. <https://doi.org/10.1021/acs.est.6b00808>.
- Deutsches Institut für Normung (DIN), 2012. Treatment of water of swimming pools and baths – Part 1: General requirements. DIN 19643.
- Dougan, D.S., Damiano, L., 2004. CO<sub>2</sub>-based demand control ventilation: Do risks outweigh potential rewards? *ASHRAE J.* 46 (10) 47–50, 52–54.
- Font-Ribera, L., Gràcia-Lavedan, E., Aragonés, N., et al., 2018. Long-term exposure to trihalomethanes in drinking water and breast cancer in the Spanish multicase-control study on cancer (MCC-SPAIN). *Environ. Int.* 112, 227–234. <https://doi.org/10.1016/j.envint.2017.12.031>.
- Gabriel, M.F., Felgueiras, F., Mourao, Z., Fernandes, E.O., 2019. Assessment of the air quality in 20 public indoor swimming pools located in the Northern Region of Portugal. *Environ. Int.* 133 (Pt B), 105274. <https://doi.org/10.1016/j.envint.2019.105274>.
- Gouveia, P., Felgueiras, F., Mourao, Z., Fernandes, E.O., Moreira, A., Gabriel, M.F., 2019. Predicting health risk from exposure to trihalomethanes in an Olympic-size indoor swimming pool among elite swimmers and coaches. *J. Toxicol. Environ. Health A* 82 (9), 577–590. <https://doi.org/10.1080/15287394.2019.1634383>.
- Fantuzzi, G., Righi, E., Predieri, G., Giacobazzi, P., Mastroianni, K., Aggazzotti, G., 2010. Prevalence of ocular, respiratory and cutaneous symptoms in indoor swimming pool workers and exposure to Disinfection By-Products (DBPs). *Int. J. Environ. Res. Public Health* 7 (4), 1379–1391. <https://doi.org/10.3390/ijerph7041379>.
- Hery, M., Hecht, G., Gerber, J.M., Gendre, J.C., Hubert, G., Rebuffaud, J., 1995. Exposure to chloramines in the atmosphere of indoor swimming pools. *Ann. Occup. Hyg.* 39 (4), 427–439. [https://doi.org/10.1016/0003-4878\(95\)00013-5](https://doi.org/10.1016/0003-4878(95)00013-5).
- Hsu, H.T., Chen, M.J., Lin, C.H., Chou, W.S., Chen, J.H., 2009. Chloroform in indoor swimming-pool air: Monitoring and modeling coupled with the effects of environmental conditions and occupant activities. *Water Res.* 43 (15), 3693–3704. <https://doi.org/10.1016/j.watres.2009.05.032>.
- International Organization for Standardization, 2000. ISO 16017-1:2000(E). Indoor, ambient and workplace air — Sampling and analysis of volatile organic compounds by sorbent tube/thermal desorption/capillary gas chromatography.
- Jacobs, J.H., Spaan, S., van Rooy, G.B.G.J., et al., 2007. Exposure to trichloramine and respiratory symptoms in indoor swimming pool workers. *Eur. Respir. J.* 29 (4), 690. <http://erj.ersjournals.com/content/29/4/690.abstract>.
- Kampel, W., Carlucci, S., Aas, B., Bruland, A., 2016. A proposal of energy performance indicators for a reliable benchmark of swimming facilities. *Energy Build.* 129, 186–198. <https://doi.org/10.1016/j.enbuild.2016.07.033>.
- Lévesque, B., Duchesne, J.-F., Gingras, S., et al., 2006. The determinants of prevalence of health complaints among young competitive swimmers. *Int. Arch. Occup. Environ. Health* 80 (1), 32–39. <https://doi.org/10.1007/s00420-006-0100-0>.
- Liu, Z., Ma, S., Cao, G., Meng, C., He, B.-J., 2018. Distribution characteristics, growth, reproduction and transmission modes and control strategies for microbial contamination in HVAC systems: a literature review. *Energy Build.* 177, 77–95. <https://doi.org/10.1016/j.enbuild.2018.07.050>.
- Lu, C.Y., Lin, J.M., Chen, Y.Y., Chen, Y.C., 2015. Building-related symptoms among office employees associated with indoor carbon dioxide and total volatile organic compounds. *Int. J. Environ. Res. Public Health* 12 (6), 5833–5845. <https://doi.org/10.3390/ijerph120605833>.
- Luoma, M., Batterman, S.A., 2000. Autocorrelation and variability of indoor air quality measurements. *AIHAJ – Am. Ind. Hyg. Assoc. J.* 61 (5), 658–668. <https://doi.org/10.1080/15298660008984575>.
- Markes Int, 2019a. ACTI-VOC low-flow sampling pump. <https://www.markes.com/Products/Sampling-accessories/Sampling/ACTI-VOC-low-flow-pump.aspx>.
- Markes Int, 2019b. Stainless steel thermal desorption sorbent tubes. <https://www.markes.com/Products/Sampling-accessories/Sorbent-tubes/Stainless-steel.aspx>.
- Andersson, M., Backman, H., Nordberg, G., et al., 2018. Early life swimming pool exposure and asthma onset in children – a case-control study. *Environ. Health* 17 (1), 34. <https://doi.org/10.1186/s12940-018-0383-0>.
- Lebon, M., Fellouah, H., Galanis, N., Limane, A., Guerfala, N., 2017. Numerical analysis and field measurements of the airflow patterns and thermal comfort in an indoor swimming pool: a case study. *Energy Effic.* 10 (3), 527–548. <https://doi.org/10.1007/s12053-016-9469-0>.
- Ministry of Culture, 2016. Idrettshaller Planlegging og bygging (Sport Facilities – Planning and Building). [https://www.regjeringen.no/contentassets/dee978d794694506bba23a57d8a76ea8/v-0989b\\_idrettshaller\\_planlegging\\_og\\_bygging\\_2016.pdf](https://www.regjeringen.no/contentassets/dee978d794694506bba23a57d8a76ea8/v-0989b_idrettshaller_planlegging_og_bygging_2016.pdf).
- Nitter, T.B., Carlucci, S., Olsen, S.N., Svendsen, K.V.H., 2019. Energy use and perceived health in indoor swimming pool facilities. *IOP Conf. Ser.* 609, 042051. <https://doi.org/10.1088/1757-899x/609/4/042051>.
- Nitter, T.B., Wolfgang, K., Svendsen Kv, H., Aas, B., 2018. Comparison of trihalomethanes in the air of two indoor swimming pool facilities using different type of chlorination and different types of water. *Water Sci. Technol.-W Sup.* 18 (4), 1350–1356. <http://www.wisaponline.com/content/early/2017/10/09/ws.2017.201.abstract>.
- Nitter, T.B., Svendsen, K.V.H., 2019b. Modelling the concentration of chloroform in the air of a Norwegian swimming pool facility—a repeated measures study. *Sci. Total Environ.* 664, 1039–1044. <https://doi.org/10.1016/j.scitotenv.2019.02.113>.
- Nitter, T.B., Svendsen, K.V.H., 2019a. UV treatment and air quality in a pool facility. *Water Sci. Technol.* 80 (3), 499–506. <https://doi.org/10.2166/wst.2019.291>.
- Norbäck, D., Björnsson, E., Janson, C., Widström, J., Boman, G., 1995. Asthmatic symptoms and volatile organic compounds, formaldehyde, and carbon dioxide in dwellings. *Occup. Environ. Med.* 52 (6), 388–395. <https://doi.org/10.1136/oem.52.6.388>.
- Norwegian Ministry of Health, 1996. Regulations for swimming facilities, swimming pools and sauna. <https://lovdata.no/dokument/SF/forskrift/1996-06-13-592>.
- Padhi, R.K., Sowmya, M., Mohanty, A.K., Bramha, S.N., Satpathy, K.K., 2012. Formation and speciation characteristics of brominated trihalomethanes in seawater chlorination. *Water Environ. Res.* 84, 2003–2009. <https://doi.org/10.2175/106143012X13415215906735>.
- Rodríguez, A., Tajuelo, M., Rodríguez, D., Seseña, S., Ruiz, P., Palop, M.L., 2018. Assessment of chemical and microbiological parameters of indoor swimming pool atmosphere using multiple comparisons. *Indoor Air* 28 (5), 676–688. <https://doi.org/10.1111/ina.12477>.
- Seppänen, O.A., Fisk, W.J., Mendell, M.J., 1999. Association of ventilation rates and CO<sub>2</sub> concentrations with health and other human responses in commercial and institutional buildings. *Indoor Air* 9 (4), 226–252. <https://doi.org/10.1111/j.1600-0668.1999.00003.x>.
- Qiao, S., Qu, Y., Ma, Y., et al., 2019. A Sensitive quantum dioxide sensor based on photo-acoustic spectroscopy with a fixed wavelength quantum cascade laser. *Sensors* 19 (9). <https://doi.org/10.3390/s1919187>.
- SINTEF Byggeforsk, 2003. Ventilasjon og avfukting i svømmehaller og rom med svømmebasseng. <https://bks.byggeforsk.no/DocumentView.aspx?documentId=534&sectionId=2>.
- United States Environmental Protection Agency, 1999. Compendium of Methods for the Determination of Toxic Organic Compounds in Ambient Air, second ed. United States Environmental Protection Agency.
- Verein Deutscher Ingenieure, 2010. Building Services in swimming baths Indoor pools (VDI 2089).
- World Health Organization, 2006. Guidelines for safe recreational water environments volume 2 swimming pools and similar environments. [http://www.who.int/water\\_sanitation\\_health/bathing/srwe2full.pdf](http://www.who.int/water_sanitation_health/bathing/srwe2full.pdf).
- World Health Organization, 2017. Guidelines for Drinking-water Quality, fourth ed. World Health Organization, Geneva, Switzerland.